1	Title: A comparison of breeding population estimators using nest and brood monitoring data
2	
3	Running title: Estimating breeding population size of birds
4	
5	Word Count: <mark>6,970</mark>
6	
7	Authors: David M. Baasch <sup>1</sup> , Trevor J. Hefley <sup>1</sup> , and Staci D. Cahis <sup>1</sup>
8	<sup>1</sup> Headwaters Corporation, 4111 4 <sup>th</sup> Avenue, Suite 6, Kearney, Nebraska 68845, USA.
9	
10	Correspondence author: David M. Baasch; (308) 390-0456; baaschd@headwaterscorp.com

#### 11 Abstract

12 1. For many species, breeding population size is an important metric for assessing population 13 status. A variety of simple methods are often used to estimate this metric for ground-nesting 14 birds that nest in open habitats (e.g., beaches, riverine sandbars). The error and bias associated 15 with estimates derived using these methods vary in relation to differing monitoring intensities 16 and detection rates. However, these errors and biases are often difficult to obtain, poorly 17 understood and largely unreported.

2. A method was developed to estimate the number of breeding pairs using counts of nests and broods from monitoring data where multiple surveys were made throughout a single breeding season (breeding pair estimator; BPE). The BPE method was compared to two commonly used estimation methods using simulated data from an individual based model that allowed for the comparison of biases and accuracy.

3. The BPE method underestimated the number of breeding pairs, but generally performed better
than the other two commonly used methods when detection rates were low and monitoring
frequency high. As detection rates and time between surveys increased, the maximum nest and
brood count method performs equally to the BPE.

4. The BPE was compared to four other methods used to estimate breeding pairs for empirically
derived data sets on the Platte River. While the BPE was not significantly different than the other
methods, it proved to provide reasonable estimates that were near the median of the other
estimators combined.

5. When data from multiple nest and brood surveys are available, the BPE appears to result in
reasonable estimates of numbers of breeding pairs. When survey data exceeds 14 days, the
maximum nest and brood count method was statistically indifferent from the BPE. Regardless of

- 34 the estimation method, investigators are encouraged to acknowledge whether the method
- 35 employed is likely to over or underestimate breeding pairs. This paper provides a means to
- 36 address that uncertainty.
- 37 Key Words: BPE, breeding pair, breeding population size, breeding population estimator,
- 38 Charadrius melodus, ground-nesting birds, interior least tern, piping plover, Platte River,
- 39 Sternula antillarum athalassos, threatened and endangered species.

## 40 Introduction

41 For threatened or endangered birds, breeding population size is an important metric for assessing recovery of the species. If the method(s) used to estimate the size of breeding 42 43 populations are not well documented, population estimates may be dissimilar and not comparable across subpopulations or within a single population over time. For example, a review 44 45 of several recovery plans, biological opinions, monitoring protocols, and reports focused on endangered interior least terns (Sternula antillarum athalassos; least tern) and threatened piping 46 plovers (Charadrius melodus) found most of these documents recommend estimating the 47 48 numbers of breeding pairs within localized areas where nesting occurs (hereafter "subpopulations"). In these documents, methods for estimating the number of breeding pairs in 49 the subpopulations included a range of methods, but no specific recommendations (Hecht and 50 51 Melvin 2009; Environment Canada 2013; Shaffer et al. 2013); included multiple methods to be employed within or between nesting seasons and therefore may not be comparable across nesting 52 seasons (Platte River Recovery Implementation Program [Program] 2011; Frost 2013, Shaffer et 53 54 al. 2013); or, in a large number of cases, were not defined and left to be chosen by the investigator (U.S. Fish and Wildlife Service [USFWS] 1988, 1989, 1990, 1996, 2003, 2006; U.S. 55 56 Army Corps of Engineers [USACE] 1993, 1999; Whitfield et al. 1996; Lutey 2002; Boettcher et al. 2007). Recovery plans for other ground-nesting bird species may suffer from similar 57 ambiguities. 58 59 The review revealed that the methods most commonly used to estimate breeding pairs

included: maximum annual adult count / two; adult count during a single standardized survey /
two (e.g., mid-June); numbers of active nest and broods observed during a single survey; and

total numbers of nests observed (Burger 1984, 1988; USACE 1993; Environment Canada 2006,

Program 2011; USFWS 2011; Frost 2013; Shaffer et al. 2013). To produce reliable estimates of breeding pairs, each of these methods require implicit assumptions. However, these assumptions may not be appropriate given the monitoring data and associated data collection protocols. As a result, comparisons of breeding pair estimates between subpopulations or through time can be unreliable and potentially misleading when the assumptions of the methods are not met. As a result, evaluations of recovery status (e.g., the number of breeding pairs in a subpopulation) using these methods can be misleading.

To date, development and evaluation of methods for estimating the number of least tern 70 71 and piping plover breeding pairs in a subpopulation has been lacking. This study focused on 72 development and evaluation of a method that uses nest and brood monitoring data, which many monitoring programs record as a normal part of monitoring efforts. The objective of our study 73 74 was to describe and evaluate a method (hereafter breeding pair estimator; BPE) for estimating breeding population size using nest and brood monitoring data. The resulting BPE method is 75 described in detail. The performance of the BPE is then evaluated against other commonly used 76 77 methods using real and simulated data.

### 78 Methods

#### 79 DATA REQURIMENTS FOR BREEDING PAIR ESTIMATOR

Our BPE assumes the number of active nests n(t) and broods b(t) within the population is known or estimable at any given time (t) during the nesting and brood rearing season (T; using parenthetical indexing notation to represent continuous time). Such data can be obtained using a variety of survey techniques such as distant observations, aerial surveys, and grid searching. Ideally, the survey technique would be able to determine the number of active nests and broods within the system on a near continuous basis. In reality, these data are typically collected at

86	discrete points in time (i.e., $t = 1, 2$ ) where it can only be assumed to approximate the						
87	continuous process. Consequently, the precise date and time nests and broods are initiated, hatch,						
88	fail, or fledge is rarely known. Therefore the time when transitions in $n(t)$ and $b(t)$ occur is						
89	unknown. In order to transform the observed discrete data into reasonable approximations of the						
90	continuous process, the following six assumptions are used to determine the date events						
91	occurred:						
92	1) The initiation date of successful nests (i.e., $\geq 1$ egg hatched) was calculated using the						
93	maximum between the period the nest was observed to be active and a known amou						
94	of time that must pass between when a nest was initiated and when $\geq 1$ egg hatched						
95	(hereafter referred to as the nest interval). A reasonable estimate of the nest interval						
96	can be obtained from the literature or from auxiliary data (e.g., band resightings).						
97	2) The initiation date of failed nests was assumed to have occurred on the date the nest						
98	was first observed. Nest and brood monitoring data do not contain information that						
99	would allow for a meaningful calculation of the nest initiation date. As such, nests						
100	with a final fate of failed or unknown were assumed to be initiated on the day they						
101	were first observed.						
102	3) Nest or brood hatching, failure, or fledging events that occurred between surveys						
103	were assumed to have occurred at the midpoint between visits. By using the midpoin						
104	between successive observations, the timing of each event was overestimated and						
105	underestimated with equal chance (Mayfield 1961; Johnson 1979; Schroeder 1997).						
106	4) The date $\geq 1$ chick fledged from of a brood was calculated using a known amount of						
107	time that must pass between when a nest hatched $\geq 1$ egg and when $\geq 1$ chick fledged						

108	(hereafter referred to as the brood interval). Reasonable estimates of the brood
109	interval could be obtained from the literature.

- 5) The minimum amount of time that must pass before a breeding pair with a failed nest
  or brood can initiate another nest was known (hereafter referred to as the renest
  interval). The renest interval can be determined from the literature or from auxiliary
- 113 data (e.g., band resightings).
- 114 6) The minimum amount of time that must pass before a breeding pair that fledges a
- brood can initiate another nest was known (hereafter referred to as the post-fledge
- 116 interval). This can be determined from the literature or auxiliary data. For species that
- 117 produce only one brood per seasons (e.g., least terns), the post-fledge interval will be
- 118 the time period from when the brood fledges until the end of the nesting season.
- 119 A visual example of the requisite data is provided (Fig. 1).

### 120 BREEDING PAIRS ESTIMATOR

Using the data and assumptions described above, breeding pair estimates were based on the sum of active nests and broods and failed nests and broods with renest intervals that extend through time *t*, and hatched broods with post-fledge interval extending through time *t* for each day of the nesting season (i.e., the assumed time step is 1 day; Fig. 1). Numbers of breeding pairs were calculated using the estimator

$$\widehat{N} = \max_{t \in T} \{ n(t) + b(t) + r(t) + f(t) \}$$
(eqn 1)

where  $\hat{N}$  is the estimated number of breeding pairs, n(t) is the number of active nests, and b(t)is the number of broods on the  $t^{th}$  day. The r(t) is the number of failed nests or broods with renest intervals extended thought the  $t^{th}$  day and f(t) is the number of fledged broods with postfledge intervals extending through the  $t^{th}$  day. The notation  $t \in S$  simply states the  $t^{th}$  day occurs 131 "within" the nesting and brood rearing season T. This estimator assumes n(t) and b(t), and by extension r(t) and f(t), are known without error, which means the number of nests and broods 132 133 counted during any given survey period can reasonably be assumed to be a census (see Simulation Experiment below for a test of this assumption). Annual estimates of breeding pairs 134 are obtained by identifying the maximum of n(t) + b(t) + r(t) + f(t) for any given day 135 during the nesting and brood rearing season (Fig. 1). To assist users, a tutorial and an excel 136 spreadsheet are provided to assist in implementation of the BPE method (Appendices S1–S2). 137 138 ALTERNATIVE BREEDING PAIR ESTIMATORS

One method commonly used to estimate the number of breeding pairs is maximum number of active nests  $n_i$  and broods  $b_i$  on any given survey (*i*; hereafter referred to as max nest and brood counts):

142

$$\widehat{N} = \max_{i \in S} \{n_i + b_i\}.$$
 (eqn 2)

Subscript indexing notation is used to represent discrete surveys. The notation  $s \in S$  states the  $i^{th}$ survey occurs "within" the discrete nesting and brood monitoring S (i.e., i = 1, 2, ..., s; where s is the total number of surveys). This method does not require "continuous" data and does not require the identity of nests or broods be uniquely identified.

147 Another commonly used estimation methods is cumulative nest counts

148 
$$\widehat{N} = \sum_{i=1}^{s} \Delta n_i, \qquad (\text{eqn 3})$$

149 where  $\Delta n_i$  is the number of new nests added during the *i*<sup>th</sup> survey (except for the first survey

150  $\Delta n_1$  is the number of nest observed). This method does not require "continuous" data, but does

151 require nests be uniquely identified.

### 152 SIMULATION EXPERMENT

153 TernCOLONY is an individual based simulation model that was developed to better 154 understand how reservoir operations and management activities affect least tern breeding 155 populations on large river systems (Lott et al. 2012 and 2013). TernCOLONY is ideal for 156 evaluating estimation methods because the model is process based, realistic, detailed, and the "true" number of breeding pairs is known. Output from 600 individual TernCOLONY simulation 157 158 runs was used to test the ability of the three methods (BPE, max nest and brood counts, and 159 cumulative nest counts) to estimate the known number of breeding pair from each model run. Each simulation included a total of 446 adults, but arrival and departure dates of individual 160 161 adults varied as did the number of adults forming breeding pairs. As a result, the number of 162 adults was the same across all simulations, but the number of breeding pairs was variable, influenced by annual habitat conditions, and was based on the number of females that initiated 163 164  $\geq$ 1 nest within the model run. In TernCOLONY, the nest period, brood period, and renest interval were variable and had a mean of 21 days, 20 days, and five days, respectively. Renesting 165 did not occur after a female produced a successful brood (fledged  $\geq 1$  chick) in TernCOLONY. 166 167 The 600 model runs incorporated multiple combinations of nesting conditions (excellent habitat with low predation or degraded habitat with high predation) and water year (high flow, 168 169 low flow, or mid-season flood) and included 30 replicates for each of the following scenarios: 1) 2 years when habitat was degraded (old), flows were high, and predation was high; 170 2) 4 years when habitat was degraded, flows were low, and predation was high; 171 172 3) 4 years when habitat was degraded, a mid-season flood occurred, and predation was high; 4) 2 years when habitat was excellent (new), flows were high, and predation was low; 173 5) 4 years when habitat was excellent, flows were low, and predation was low; 174 175 6) 4 years when habitat was excellent, a mid-season flood occurred, and predation was low.

176 The BPE (eqn. 1), maximum number of active nests and broods (eqn. 2), and cumulative 177 number of nests (eqn. 3) all assume the number of nests or broods can be detected perfectly. The 178 assumption of perfect detection is unrealistic. Because all estimation methods are sensitive to this 179 assumption, a binomial distribution was used to simulate non-detection of nests and broods. In 180 addition, estimates from each method are sensitive to sampling interval (i.e., how frequently data 181 are collected). Each model run was sampled every third, seventh, and fourteenth day and once 182 during the season (June 15) assuming a detection probability of 0.50, 0.75 and 1.00. This data was then used to estimate breeding pairs using the BPE (eqn. 1), maximum number of active 183 184 nests and broods (eqn. 2), and cumulative number of nests (eqn. 3).

The BPE model assumptions included a nest interval of 21 days, a brood interval of 20 days, a renest interval of five days and a post-fledge interval extending to the end of the nesting season (i.e., renesting did not occur after producing a successful brood). Results of the BPE are presented as  $\hat{N}$  divided by the known number of breeding pairs for each model run with all scenarios combined (see Fig. 2). Ratios of 1.00 represents a perfect estimate of the known number of breeding pair and values above or below 1.00 indicate over or under estimates of breeding pairs, respectively.

#### 192 Case Study

#### 193 BACKGROUND

The case study used data from the Associated Habitat Reach (AHR) of the central Platte River Valley beginning at the junction of U.S. Highway 283 and Interstate 80 near Lexington, Nebraska, and extending eastward to Chapman, Nebraska, USA (Program 2006, 2011). The AHR provides breeding habitat for a variety of shorebirds, including the federally endangered least tern and threatened piping plover (Faanes 1983; Sidle and Kirsch 1993; Jenniges and

Plettner 2008). Throughout their range, least terns and piping plovers nest sympatrically on inchannel (sandbars), off-channel (sand and gravel mines), and shoreline nesting habitats (Ziewitz
et al. 1992; Jenniges and Plettner 2008).

The study area represents a sub-population of least terns and piping plovers that occur along the central Platte River in Nebraska. Many areas within these species' ranges are surveyed to count and monitor nests and broods which results in data similar to data collected in the AHR. Throughout the species' range, at least eight methods are used to calculate numbers of breeding pairs (USACE 1993; Program 2011; USFWS 2011; Frost 2013; Shaffer et al. 2013). At the moment, it is unclear how reported counts using such disparate methods can be reconciled to determine the status of the breeding populations.

#### 209 FIELD SURVEY TECHNIQUES

210 The least tern and piping plover monitoring protocol implemented in the AHR from 211 2001–2014 was comprised of two main components: 1) semi-monthly river surveys and 2) semi-212 monthly surveys of historic, existing, and potential sandpit nesting sites within the AHR 213 (Program 2011). During these surveys, numbers of adults, nests, and chicks of each species 214 observed were recorded. Nests and broods located during surveys were monitored at least twice 215 per week as long as nests or broods were present and new nests and broods were located during 216 each survey. The frequency of survey and monitoring efforts (twice weekly) allowed detection of a large, but unknown proportion of nests within the AHR and allowed the derivation of fairly 217 218 accurate estimates of the timing of nest or brood failures as well as hatching and fledging events. 219 The data required to estimate the number of breeding pairs using BPE along with calculations 220 used in the BPE are available in a spreadsheet archived on the Dryad Digital Repository (see 221 Data Accessibility; Appendix S2).

## 222 BREEDING PAIR ESTIMATE AND COMPARISON

223 Monitoring data collected in the AHR was used to compare five methods of estimating 224 breeding pairs annually: BPE (eqn. 1), cumulative nest counts (eqn. 3), maximum number of 225 nests and broods observed during mid-month and semi-monthly surveys (eqn. 2), number of nests and broods observed on 15 June (eqn. 2 with a single sample period), and half of the 226 227 maximum number of adults observed during mid-month and semi-monthly surveys of the AHR. The last method (half of the maximum number of adults) was included because it is a common 228 229 method used to estimate the number of breeding pairs in the study area and is currently used for 230 other subpopulations. We define 'nesting period' as the time a nest was first initiated (first egg in the scrape) to the time when the nest hatched. We are fully aware the 'nesting period' could be 231 as much as 24 - 26 days from when a nest is initiated to when it hatches, however, our goal was 232 233 to develop a method that was conservative, but yet a reasonable estimate of the number of 234 breeding pair in the AHR.

235 Annual least tern and piping plover breeding pair counts were estimated using eqn. 1, which required calculations of n(t), b(t), r(t), and f(t) for each species and day of the nesting 236 seasons. The BPE assumptions for least terns included a nest interval of 21 days (incubation 237 238 period), a brood interval of 21 days, a renest interval of five days and a post-fledge interval that 239 extended to the end of the nesting season (i.e., no renesting after successfully fledging a brood). 240 The renest interval of five days was based on band-resight data, observations of nesting 241 chronology, and published data (Massey and Fancher 1989; Lingle 1990 and 1993; Lott et al. 242 2012; Program unpublished data). The BPE assumptions for piping plover included a nest 243 interval of 28 days, a brood interval of 28 days, a renest interval of five days, and a post-fledge 244 interval of five days. The renest and post-fledge intervals were based on band-resight data,

observations of nesting chronology, and published data (Roudybush et al. 1979; Amat et al.
1999; Shaffer et al. 2013; Program 2014).

An important goal of the Program monitoring protocol is to detect population trends. Simple linear regression was used to detect trends in the time series of 2001–2014 data based on the breeding pair estimates. Regression coefficient and associated 95% CIs were reported. A pair-wise correlation matrix was also developed for each estimation method for comparison purposes.

252 **Results** 

## 253 SIMULATION EXPERIMENT

254 The BPE and maximum nest and brood count methods usually resulted in statistically indistinguishable breeding pair estimates negatively biased (underestimated) under all sampling 255 256 intensities and detection rates except for the 3-day sampling with perfect detection. The magnitude of the negative bias depended on the sampling interval and detection rates (Fig. 2). 257 The cumulative number of nests method typically overestimated the number of breeding pairs 258 259 when sampling occurred frequently (3-day and 7-day) and underestimated when sampling 260 occurred less frequently. As with the other methods, the magnitude of the bias depended on the 261 detection rate (Fig. 2). When detection was low and only a single mid-June survey was simulated, estimates of the known breeding pair count were severely underestimated (negatively 262 biased) regardless of the estimation method. This result was not unexpected as it would be highly 263 264 unlikely that all nests and broods would be present during any single survey date.

Of the three methods tested, the BPE was influenced the least by detection rates. The BPE was most sensitive to sampling interval when detection was low (i.e., 50%) and estimates improved as detection increased to 100%. When detection was high and the sampling interval

was short (i.e., 3-day sampling interval), the BPE resulted in an average breeding pair estimate that was 18% (range = 16% - 21%) less than the true value. By design, estimates from the BPE and maximum nest and brood count methods were identical when only a mid-June survey was simulated.

Estimates of breeding pair counts derived using the maximum nest and brood count method were also most often underestimated. The maximum nest and brood count method was the least influenced by sampling intensity of all methods tested (Fig. 2). Results of the maximum nest and brood count method were indistinguishable from the BPE when detection was assumed to perfect. When detection was low, this method typically resulted in the most negatively biased estimates (underestimated) of all methods tested (54% to 71% low).

The cumulative nest count method produced breeding pair counts that ranged from highly 278 279 overestimated (+53%) to highly underestimated (-72%). Results of this method were highly 280 dependent on the survey interval and detection rate. Estimates obtained from cumulative nest counts were most exaggerated (overestimated) when the sampling interval was short and 281 282 detection was high and declined as the sampling interval increased and detection decreased. 283 When detection was perfect, the cumulative nest count method overestimated the known 284 breeding pair counts by 24% - 53% when multiple surveys were implemented. When detection was perfect and only a single sampling interval was used to obtain estimates, breeding pair 285 counts were underestimated (-43%). 286

287 CASE STUDY

Trends in AHR least tern breeding pair estimates were positively correlated and tended to follow a similar increasing pattern for all nest and brood monitoring methods tested (Table 1; Fig. 3). Regression coefficients for the trend line associated with each method varied from 1.35

291 (Adult count/2) to 5.55 (cumulative nest counts). The 95% CIs for all trend lines, however, 292 overlapped indicating the regression coefficient for all five methods could be the same (Table 1). 293 As with the simulation experiment, least tern 15-June nest and brood counts provided the lowest 294 estimate of breeding pairs. Maximum nest and brood counts obtained from mid-month (2001– 2009) and semi-monthly (2010–2014) surveys were highly correlated with BPE (r=0.96). 295 296 Cumulative nest counts generally provided the highest annual estimates of breeding pairs. 297 However, it is known that this method would always be biased high unless all breeding pairs only produced a single nest each year. 298

299 Similar to least terns, trends in piping plover breeding pair estimates tended to follow a 300 similar increasing pattern for all methods tested (Table 1; Fig. 4). Regression coefficients for the trend line associated with each method varied from 1.24 (15 June nest and brood counts) to 1.97 301 302 (cumulative nest counts). The 95% CIs for all trend lines overlapped indicating the regression 303 coefficient for all five methods could be the same (Table 1). Adult piping plover counts tended to 304 be most comparable to breeding pair estimates generated by the BPE. The 15-June nest and 305 brood count and maximum mid-month and semi-monthly methods for piping plovers resulted in 306 similar estimates; however, these methods were at times up to 47% lower than the BPE for 307 estimating breeding pairs. The cumulative nest count method provided the highest annual estimates of breeding pairs and at times was 53% (range 10% - 53%) higher than the BPE. 308

309 Discussion

310 SIMULATED EXPERIMENT

In TernCOLONY simulations, nest and brood counts were perfectly observable in nearly continuous time with no error. The performance of three breeding pair estimation methods was compared by sampling from the simulated populations. This allowed for comparison of bias 314 across methods relative to perfect knowledge of the population. Many papers have addressed the 315 ubiquitous problem of imperfect detection in wildlife surveys (Thompson 2002; Lott 2006). In 316 this analysis, detection rates were varied from 50% to perfect detection (100%) and sampling 317 interval from a single survey to a 3 day sampling interval. Length of the interval between 318 sampling periods to a given nesting area can bias detection toward successful nests, potentially 319 leading to underestimates of initiated nests and nest loss rate and an inability to quantify causes 320 of nest loss (Shaffer et al. 2013). Incorporating detection rates and sampling interval into the analysis allowed quantification of the sensitivity of breeding pair estimation methods to these 321 322 known issues.

The BPE is a method developed by the Program to estimate the number of breeding pairs 323 using nest and brood monitoring data that includes a rest period between lost nests or broods and 324 325 renesting by an individual pair. Though the method employed by Shaffer et al. (2013) was similar to the BPE that has been used in the AHR for several years, the minimum breeding 326 327 population (MINBPOP) method does not account for breeding pairs that renest after losing a nest 328 or brood. Thus, the implication of this is that results of the MINBPOP method are identical to the 329 cumulative nest count method used in our study as every nest counted was assumed to be 330 associated with a unique breeding pair. We feel the BPE will be most useful for breeding bird populations that nest in open habitats (e.g., sandbars, beaches, etc.) for which numbers of nests 331 and broods counted on any given sampling period can reasonably be assumed to be high, but less 332 333 than perfect (i.e., detection <100%). Results from the simulation study show the BPE tended to produce the most unbiased and least variable estimate of the total number of breeding pairs in a 334 335 population so long as sampling occurred fairly frequently (i.e., 3-day interval) and detection was 336 assumed to be imperfect. Though variable, estimates across individual nest and brood monitoring

methods at the similar levels of detection and sampling interval tested were statisticallyindistinguishable.

339 In many cases, the primary goal may be to estimate the number of breeding pairs in the 340 population. Another goal may be the development of an index of breeding pair abundance that is 341 comparable across different study areas and sampling intensities or designs. For example, the 342 sampling intensity (e.g., 3-day, 7-day, etc.) may vary over time due to availability of funding within a study area. If the goal was to produce an index that is comparable when sampling 343 interval is variable, then the maximum nest and brood count method appears to be less sensitive 344 345 to a variable sampling interval; however, estimates were consistently lower than the known number of breeding pairs and estimates obtained by the BPE. 346

347 CASE STUDY

An illustrative example was provided using monitoring data for least terns and piping plovers collected in the AHR to evaluate management actions for a large scale species recovery program. Recovery plans require numbers of pairs to be estimated to determine if recovery goals have been met. If pair estimates are used to estimate trends, all five methods produced coefficient estimates that indicated the subpopulation within the AHR was increasing and, based on overlapping CIs, coefficients obtained from all breeding pair estimation methods were not statistically different.

Though recovery goals for least terns and piping plover are based on maintenance of pairs of each species in subpopulations for a predetermined time period, recovery plans provide no guidance for how pairs are to be determined. Although we evaluated multiple disparate methods of estimating breeding pairs, our analyses indicated there were no statistical differences between methods in regards to estimating trends in the population (Table 1). When comparing

360 regression coefficients for least terns, however, the maximum nest count and the adult count 361 methods resulted in estimates of slope that were more than four times greater for the maximum 362 nest count method. However, our inability to detect a difference between methods was due to the 363 high variability in counts over time.

If adult counts are to be used to determine numbers of pairs, we feel it is important to 364 365 acknowledge and attempt to account for several factors including some adults are not actively paired during the nesting season, obtaining accurate counts of adults may be difficult in large 366 colonies, assessing detection rates for adults may be difficult given their high mobility and 367 368 foraging behaviors, and similar to the cumulative nest count method, adult counts have been 369 reported to result in breeding pair estimates that range widely from overestimated to underestimated (Sherfy et al. 2012, Shaffer et al. 2013). We were not able to estimate breeding 370 371 pair counts in our simulation study and therefore cannot provide any guidance as to how this method compares to nest-based methods used in our study. However, we feel it is safe to assume 372 373 more adults would equate to more breeding pairs and thus using the adult count method to 374 estimate trends in breeding pair counts likely would result in a similar pattern as using other 375 methods (Figures 3 and 4).

In the AHR and other areas, least terns and piping plovers nest on bare sand habitat provided on in-channel sandbars and off-channel sand and gravel mines (Program 2012, 2014). Given high intensity monitoring (e.g., at least twice weekly) and characteristics of habitat used by least terns and piping plovers (bare sand), we suspect detections rates in the AHR are high and believe nest and brood counts can be assumed to approximate a census (Roche et al. 2014). If this is the case, the BPE and maximum nest and brood count methods result in estimates of breeding pairs that were indistinguishable. The assumption of perfect detection, however, should

383 be justified based on the ecology of the species studied and survey methodology employed. If 384 information about the detection process and rate for nests and broods is available (e.g., Roche et al. 2014), the BPE could easily be extended to incorporate this information. For example, one 385 386 could use the estimated probability of detection of nests and broods to adjust the number of nests 387 and broods that are active on a given day (n(t) in eqn 1). For example, assuming a detection rate 388 of 75% and given the high intensity sampling that occurs within the AHR, results of our simulation indicate estimates of breeding pairs derived using the BPE may in fact be 389 390 approximately 18% lower than reality.

391 Conclusion

392 All methods examined resulted in estimates of numbers of breeding pairs that were not significantly different, however, this was likely due to the variability in the data as difference 393 between estimates were as much as fourfold. Thus, the need a unified approach for estimating 394 395 these metrics throughout a species' range is evident. A unified approach would allow for direct 396 comparisons of breeding pair counts and productivity measures (fledge ratios, etc.) between 397 regions where a species nests, so long as the nesting and brood rearing period were defined in a 398 similar manner. When nest and brood monitoring data are collected at intervals of less than 14 399 days, the BPE provided estimates of breeding pairs that were the most precise and accurate, 400 especially when detection was assumed to be less than perfect. If survey intervals exceed 14 days 401 and detection can be assumed to be nearly perfect, the maximum nest and brood count method 402 results in estimates that were generally conservative (underestimate breeding pairs), but 403 indistinguishable from estimates produced by the BPE. The cumulative nest count method is 404 highly sensitive to monitoring intervals and results in breeding pair estimates that range from 405 highly under estimated to highly over estimated. We recommend practitioners refer to the

simulation portion of this analysis to develop a better understanding of the sensitivity of
currently used estimators to monitoring frequency and detection and whether or not those
estimators are likely to over or under estimate breeding pairs. We also recommend researchers
enter nest and brood monitoring data into a standardized database, such as Appendix S2, so

- 410 comparable assumptions and estimates can be derived throughout the study species' range.
- 411 Acknowledgements

We would like to thank the many technicians and biologists that assisted with the collection of data used in the case study. We would also like to thank all the programmers and individuals who assisted with the development of TernCOLONY; especially Casey Lott and Colin Sheppard who provided us the raw data. The Platte River Recovery Implementation Program provided funding for this work.

417

# 418 Data accessibility

The source data for the tern and plover estimation exercises are archived in the Dryad DigitalRepository doi:XXX.

421

## 422 Literature Cited

- Amat, J.A., Fraga, R.M. & Arroyo, G.M. (1999) Replacement clutches by Kentish plovers: The
  Condor, 101, 746-751.
- 425 Baasch, D.M. (2014) Platte River Recovery Implementation Program: 2012-2013 interior least
- 426 tern and piping plover monitoring and research report for the central Platte River,
- 427 Nebraska. URL
- 428 https://www.platteriverprogram.org/PubsAndData/ProgramLibrary/PRRIP%202012-

- 429 2013%20Tern%20and%20Plover%20Monitoring%20and%20Research%20Report.pdf
- 430 [Accessed 19 November 2014].
- 431 Boettcher, R., Penn, T., Cross, R.R., Terwilliger, K.T. & Beck, R.A. (2007) An Overview of the
- 432 Status and Distribution of Piping Plovers in Virginia. Waterbirds, 30, 138-151.
- 433 Burger, J. (1984) Colony stability in least terns. Condor, 86, 61-67.
- Burger, J. (1988) Social attraction in nesting least terns: effects of numbers, spacing, and pair
  bonds. Condor, 90, 575-582.
- 436 Environment Canada (2006) Final Recovery Strategy for the Piping Plover (Charadrius melodus
- 437 *circumcinctus*) in Canada. *Species at Risk Act* Recovery Strategy Series. Environment
- 438 Canada, Ottawa. vi + 30 pp. URL <u>http://www.registrelep-</u>
- 439 <u>sararegistry.gc.ca/default.asp?lang=En&n=B5D30A52-1</u> [Accessed 19 November 2014].
- 440 Environment Canada (2013) Action Plan for the Piping Plover (*Charadrius melodus*
- 441 *circumcinctus*) in Ontario. Species at Risk Act Action Plan Series. Environment Canada,
- 442 Ottawa. URL <u>http://www.registrelep-</u>
- 443 <u>sararegistry.gc.ca/virtual\_sara/files/plans/ap\_pluvier\_siffleur\_piping\_plover\_circumcinct</u>
- 444 <u>us\_ontario\_0213\_e.pdf</u> [Accessed 19 November 2014].
- 445 Faanes, C.A. (1983) Aspects of the nesting ecology of least terns and piping plovers in central
- 446 Nebraska. Prairie Naturalist, 15, 145-154.
- 447 Frost, N. (2013) California least tern breeding survey: 2012 season. URL
- 448 https://nrm.dfg.ca.gov/FileHandler.ashx?DocumentID=75039 [Accessed 19 November
- 449 2014].
- 450 Hecht, A. & Melvin, S.M. (2009) Population Trends of Atlantic Coast Piping Plovers, 1986–
- 451 2006. Waterbirds, 32, 64-72.

- Jenniges, J.J. & Plettner, R.G. (2008) Least Tern nesting at human created habitats in central
  Nebraska. Waterbirds, 31, 274-282.
- Johnson, D.H. (1979) Estimating nest success: the Mayfield method and an alternative. Auk, 96,
  51-661.
- 456 Lingle, G.R. (1990) Least tern and piping plover nesting ecology along the central Platte River
- 457 Valley, Nebraska. Platte River Whooping Crane Maintenance Trust, Progress Report.
- Lingle, G.R. (1993) Causes of nest failure and mortality of Least Terns and Piping Plovers along
- 459 the central Platte River. In: *Proceedings of the Missouri River and its tributaries: Piping*
- 460 Plover and Least Tern Symposium (eds K.F. Higgins & M.R. Brashier), pp. 189-191.
- 461 Brookings, SD: South Dakota State University.
- Lott, C.A. (2006) Distribution and abundance of the interior population of least tern (Sternula
- 463 *antillarum*) 2005: a review of the first comprehensive range-wide survey in the context of
- 464 historic and ongoing monitoring efforts. ERDC/EL TR-06-13. Vicksburg, MS: U.S.
- 465 Army Engineer Research and Development Center. URL
- 466 <u>http://el.erdc.usace.army.mil/elpubs/pdf/trel06-13.pdf</u> [Accessed 19 November 2014].
- Lott, C.A., Railsback, S.F. & Sheppard, C.J.R. (2012) TernCOLONY 1.0 Model Description.
- 468 URL: <u>http://el.erdc.usace.army.mil/elpubs/pdf/crel12-3.pdf</u> [Accessed 19 November
  469 2014].
- 470 Lott, C.A., Railsback, S.F., Sheppard, C.J.R. & Koohafkan, M.C. (2013) Developing and Testing
- 471 TernCOLONY 1.0: An Individual-based Model of Least Tern Reproduction. URL
- 472 <u>http://el.erdc.usace.army.mil/elpubs/pdf/crel13-2.pdf</u> [Accessed 19 November 2014].
- 473 Lutey, J.M. (2002) Species Recovery Objectives for Four Target Species in the Central and
- 474 Lower Platte River (Whooping Crane, Interior Least Tern, Piping Plover, Pallid

- 475 Sturgeon). United States Fish and Wildlife Service. URL
- 476 <u>http://cwcbweblink.state.co.us/WebLink/0/doc/169287/Page4.aspx</u> [Accessed 19
  477 November 2014].
- 478 Massey, B.W. & Fancher, J.M. (1989) Renesting by California Least Terns. Journal of Field
  479 Ornithology, 60, 350-357.
- 480 Mayfield, H. (1961) Nesting success calculated from exposure. Wilson Bulletin, 73, 255-261.
- 481 Platte River Recovery Implementation Program [Program] (2006) Final Platte River Recovery
- 482 Implementation Program Adaptive Management Plan. U.S. Department of the Interior,
- 483 State of Wyoming, State of Nebraska, State of Colorado. URL
- 484 https://www.platteriverprogram.org/PubsAndData/ProgramLibrary/PRRIP%20Attachme
- 485 <u>nt%203%20-%20adaptive\_management\_plan.pdf</u> [Accessed 9 July 2014].
- 486 Platte River Recovery Implementation Program [Program] (2011) Monitoring the Abundance,
- 487 Distribution, Reproductive Success, and Reproductive Habitat parameters of Least Terns
- 488 and Piping Plovers on the Central Platte River (2011). URL
- 489 https://www.platteriverprogram.org/PubsAndData/ProgramLibrary/PRRIP%20Tern%20a
- 490 <u>nd%20Plover%20Monitoring%20Protocol%20(2011).pdf</u> [Accessed 9 July 2014].
- 491 Platte River Recovery Implementation Program [Program] (2012) State of the Platte Report –
- 492 Executive Summary. URL
- 493 <u>https://www.platteriverprogram.org/PubsAndData/ProgramLibrary/2012%20State%20of</u>
- 494 <u>%20the%20Platte%20Executive%20Summary.pdf</u> [Accessed 9 July 2014].
- 495 Roche, E., Shaffer, T.L., Anteau, M.J., Sherfy, M.H., Stucker, J.H., Wiltermuth, M.T. &
- 496 Dovichin, C.M. (2014) Detection probability of least tern and piping plover chicks in a
- 497 large river system. The Journal of Wildlife Management, 78, 709-720.

498	Roudybush, E., Grau, C.R., Petersen, M.R., Ainley, D.G., Hirsch, K.V., Gilman, A.P. & Patten,						
499	S.M. (1979) Yolk formation in some Charadriiform birds: The Condor, 81, 293-298.						
500	Schroeder, M.A. (1997) Unusually high reproductive effort by sage grouse in a fragmented						
501	habitat in North-Central Washington. Condor, 99, 933-941.						
502	Shaffer, T.L., Sherfy, M.H., Anteau, M.J., Stucker, J.H., Sovada, M.A., Roche, E.A.,						
503	Wiltermuth, M.T., Buhl, T.K. & Dovichin, C.M. (2013) Accuracy of the Missouri River						
504	Least Tern and Piping Plover Monitoring Program–Considerations for the Future. Open						
505	File Report 2013-1176. URL <u>http://pubs.usgs.gov/of/2013/1176/pdf/of2013-1176.pdf</u>						
506	[Accessed 19 November 2014].						
507	Sherfy, M.H., Anteau, M.J., Shaffer, T.L., Sovada, M.A. & Stucker, J.H. (2012) Foraging						
508	Ecology of Least Terns and Piping Plovers Nesting on Central Platte River Sandpits and						
509	Sandbars. US Geological Survey Open-File Report 2012–1059. URL						
510	https://www.platteriverprogram.org/PubsAndData/ProgramLibrary/Foraging%20Ecology						
511	%20of%20Least%20Terns%20and%20Piping%20Plovers%20Nesting%20on%20Central						
512	%20Platte%20River%20Sandpits%20and%20Sandbars.pdf [Accessed 9 July 2014].						
513	Sidle, J.G., and Kirsch, E.M. (1993) Least tern and piping plover nesting at sand pits in						
514	Nebraska. Colonial Waterbirds, 16, 139-148.						
515	Thompson, W.L. (2002) Toward reliable bird surveys: Accounting for individuals present but						
516	not detected. Auk, 119, 18-25.						
517	U.S. Army Corps of Engineers [USACE] (1993) Guidelines for monitoring least terns and piping						

- 518 plovers. Omaha, Nebraska, U.S. Army Corps of Engineers, Omaha District, unpublished
- 519 report, 42 p.

- 520 U.S. Army Corps of Engineers [USACE] (1999) Missouri River region FY 1998 interior least
- 521 tern and piping plover biological opinion compliance report and permit activity report.
- 522 URL <u>http://www.dtic.mil/dtic/tr/fulltext/u2/a382674.pdf</u> [Accessed 9 July 2014].
- 523 U.S. Fish and Wildlife Service [USFWS] (1988) Great Lakes and Northern Great Plains Piping
- 524 Plover Recovery Plan. U.S. Fish and Wildlife Service; Twin Cities, Minnesota. URL
- 525 <u>http://www.nwk.usace.army.mil/Portals/29/docs/regulatory/ksdredging/GreatLakesRecov</u>
- 526 <u>eryPlan-May1988.pdf</u> [Accessed 19 November 2014].
- 527 U.S. Fish and Wildlife Service [USFWS] (1989) Roseate tern recovery plan-northeast
- 528 population. U.S. Fish and Wildlife Service; Newton Corner, Massachusetts. URL
- 529 <u>http://www.fws.gov/verobeach/msrppdfs/roseatetern.pdf</u> [Accessed 19 November 2014].
- 530 U.S. Fish and Wildlife Service [USFWS] (1990) Recovery plan for the interior population of the
- 531 least tern (*Sterna antillarum*). U.S. Fish and Wildlife Service, Twin Cities, MN. URL
- 532 <u>http://www.fws.gov/montanafieldoffice/Endangered\_Species/Recovery\_and\_Mgmt\_Plan</u>
- 533 <u>s/Least\_Tern\_Recovery\_Plan.pdf</u> [Accessed 11 November 2014].
- 534 U.S. Fish and Wildlife Service [USFWS] (1996) Piping plover (Charadrius melodus), Atlantic
- 535 Coast Population, Revised Recovery Plan. U.S. Fish and Wildlife Service; Hadley,
- 536 Massachusetts. URL <u>http://www.fws.gov/northeast/pipingplover/pdf/summary.pdf</u>
- 537 [Accessed 11 November 2014].
- 538 U.S. Fish and Wildlife Service [USFWS] (2003) Amendment to the 2000 Biological Opinion on
- 539 the operation of the Missouri River main stem reservoir system, operation and
- 540 maintenance of the Missouri River bank stabilization and navigation project, and
- 541 operation of the Kansas River reservoir system. URL <u>http://www.nwd-</u>
- 542 mr.usace.army.mil/mmanual/FinalBO2003.pdf [Accessed 9 July 2014].

- 543 U.S. Fish and Wildlife Service [USFWS] (2006) Biological opinion on the Platte River Recovery
   544 Implementation Program. URL
- 545 <u>https://www.platteriverprogram.org/PubsAndData/ProgramLibrary/TC-</u>
- 546 <u>R569PRRIP%20Biological%20Opinion.pdf</u>. [Accessed 9 July 2014].
- 547 U.S. Fish and Wildlife Service [USFWS] (2011) Abundance and productivity estimates 2010
- 548 update: Atlantic Coast piping plover population. Sudbury, Massachusetts. 4 pp.
- Whitfield, D.P., Brade, J.J., Burton, R.W., Hankinson, K.W. & Young, S. (1996) The abundance
  of breeding Knot *Calidris canutus islandica*. Bird Study, 43, 290-29.
- 551 Ziewitz, J.W., Sidle, J.G., and Dinan J.J. (1992) Habitat conservation for least terns and piping
- plovers on the Platte River, Nebraska. Prairie Naturalist 24, 1-20.

## 553 Supporting Information

- Additional Supporting Information may be found in the online version of this article.
- 555 Appendix S1. Tutorial for using Appendix S2 to implement Breeding Population Estimator
- 556 (BPE).
- 557 Appendix S2. Excel Spreadsheet for implementing our Breeding Population Estimator (BPE).
- 558 Appendix S3. Figure S1.



**Figure 1.** Example showing how nest and brood monitoring data and a user-defined nest interval (21 days), brood interval (21 days), renest interval (five days) and post-fledge interval were used to estimate breeding pairs. In this example the post-fledge interval extends from the time a brood fledged to the End of Breeding Season (EOBS) as the species in this hypothetical example did not renest after fledging a brood (blue bars extending to the right side of the renest interval). The grey shaded area indicates when the maximum numbers of breeding pairs (three) occurred. The vertical dashed blue lines represent a hypothetical sampling interval that occurred every 10 days. The Breeding Population Estimator (BPE) assumes sampling occurs at sufficient regularity that the maximum number of breeding pairs can reliability be estimated.









Figure 3. Five estimates of least tern breeding pairs within the central Platte River Valley (top). 575

An evaluation of how each estimate compares to estimates from our breeding pair estimator 576 (BPE; bottom). The comparison in the bottom plot was calculated as (x-BPE)/BPE, where x is 577







583 estimator (BPE; bottom). The comparison in the bottom plot was calculated as (*x*-BPE)/BPE,

584 where x is the estimate obtained using one of the four other methods.

**Table 1.** Regression coefficient and 95% confidence interval (CI) from a trend analysis of various breeding pair estimates obtained from data of least terns and piping plovers, 2001–2014. Also reported for comparison purposes is the pair-wise correlation matrix between each estimation method.

Estimator	Regression		Correlation				
	Coefficient (95% CI)	BPE	Cumulative	15 June Nests	Semi-monthly Nests	Adult Count/2	
			Nest Counts	and Broods	and Broods		
Least terns							
Breeding pair estimator (BPE)	3.24 (1.48-4.99)	1.00	0.98	0.73	0.96	0.67	
Cumulative Nest Counts	5.55 (3.13-7.98)		1.00	0.70	0.94	0.58	
15 June Nests and Broods	2.11 (0.92-3.29)			1.00	0.83	0.40	
Semi-monthly Nests and Broods	3.19 (1.73–4.66)				1.00	0.63	
Adult Count/2	1.35 (-0.69–3.38)					1.00	
Piping plovers							
Breeding pair estimator (BPE)	1.28 (0.68–1.88)	1.00	0.95	0.95	0.97	0.93	
Cumulative Nest Counts	1.97 (1.06–2.87)		1.00	0.88	0.90	0.86	
15 June Nests and Broods	1.24 (0.50–1.98)			1.00	0.99	0.85	
Semi-monthly Nests and Broods	1.35 (0.64–2.07)				1.00	0.90	
Adult Count/2	1.39 (0.61–2.17)					1.00	